



Anthropogenic fine particulate matter pollution will be exacerbated in eastern China due to 21st century GHG warming

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Abstract. China has experienced a substantial increase in severe haze events over the past several decades, which is primarily attributed to the increased pollutant emissions caused by its rapid economic development. The climate changes observed under the warming scenarios, especially those induced by increases in greenhouse gases (GHGs), are also conducive to the increase in air pollution. However, how the air pollution changes in response to the GHG warming has not been thoroughly elucidated to date. We investigate this change using the century-long large ensemble simulations with the Community Earth System Model 1 (CESM1) with the fixed anthropogenic emissions at the year 2005. Our results show that although the aerosol emission is assumed to be a constant throughout the experiment, anthropogenic air pollution presents positive responses to the GHG-induced warming. The anthropogenic PM_{2.5} concentration is estimated to increase averaged over eastern China at the end of this century, but varying from regions, with an increase over northwestern part of eastern China and a decrease over southeastern part. Similar changes can be observed for the light air pollution days. However, the severe air pollution days are reported to increase across eastern China at the end of this century, particularly around the Jing–Jin–Ji region. Further research indicates that the increased stagnation days and the decreased light precipitation days are the possible causes of the increase in PM_{2.5} concentration, as well as the anthropogenic air pollution days. Estimation shows that the effect of climate change induced by the GHG warming can account for 11 %–28 % of the changes in anthropogenic air pollution days over eastern China. Therefore, in the future, more

stringent regulations on regional air pollution emissions are needed to balance the effect from climate change.

1 Introduction

The extraordinarily rapid development of China has caused extremely high aerosol loading and gaseous pollutant emissions that have enveloped most regions across China in recent decades. The increased pollutant emissions, particularly for the particulate matter finer than 2.5 μm in aerodynamic diameter (PM_{2.5}), generally result in severe haze events and present a major threat to public health (Gao et al., 2017; Tang et al., 2017; Wang, 2018), crop production (Tie et al., 2016), and regional climates (Cao et al., 2016). For example, the annual averaged PM_{2.5} in Beijing exceeded 75 $\mu\text{g m}^{-3}$ during 2009–2016 (Fig. 1b), which is more than 3 times the recommended 24 h standard (25 $\mu\text{g m}^{-3}$) of the World Health Organization (WHO). This degeneration of the air pollution across China, which is similar to that in Beijing, is primarily caused by the integrated effects of high emissions and poor ventilation (Chen and Wang, 2015; Y. Zhang et al., 2016). Many efforts are thus underway to reduce emissions that cause severe haze pollution. However, the question remains as to whether climate change will offset or facilitate these efforts.

Recent studies have documented that the exacerbation of air quality over eastern China was partly modulated by meteorological conditions and climate variability that are generally conducive to the severe haze occurrences (Li et al.,

2018; Liao and Chang, 2014; Wang and Chen, 2016; Yang et al., 2016; Zhang et al., 2014; Z. Zhang et al., 2016). Specifically, Wang et al. (2015) revealed that the shrinking Arctic sea ice favours less cyclone activity and a more stable atmosphere conducive to haze formation, which can explain approximately 45 %–67 % of the interannual to interdecadal variability of winter haze days over eastern China. Besides Arctic sea ice, other decadal variability and changes, including a weak East Asian winter monsoon (Jeong et al., 2017; Li et al., 2016; Yin et al., 2015), strong El Niño–Southern Oscillation (Gao and Li, 2015; Zhao et al., 2018), high Pacific decadal oscillation (Zhao et al., 2016), and high Arctic oscillation (Cai et al., 2017), may have contributed. In addition, the increasing winter haze days over eastern China may also be linked to the low boundary layer height (Huang et al., 2018; Wang et al., 2018), weakened northerly winds (Yang et al., 2017a), decreased relative humidity (Ding and Liu, 2014), and increased sea surface temperature (Xiao et al., 2015; Yin and Wang, 2016; Yin et al., 2017).

Global warming generally presents an adverse impact on the haze pollution across China. Simulations of the dynamic downscaling by the regional climate model RegCM4 under the RCP4.5 (Representative Concentration Pathway) scenario have shown that the air environment carrying capacity tends to decrease, and the weak ventilation days tend to increase, in the 21st century across China, suggesting an increase in the haze pollution potential compared to the current state (Han et al., 2017). Furthermore, Cai et al. (2017) projected that the days conducive to severe haze pollution in Beijing would increase by 50 % at the end of the 21st century (2050–2099) under the RCP8.5 scenario compared to the historical period.

These qualitative estimations of the haze pollution response to climate changes are generally derived from the *potential* changes in the corresponding meteorological conditions indirectly. No studies to date quantitatively assessed the simulated PM directly. How the fine particulate matter pollution changes in response to the global warming in China has not been thoroughly elucidated to date. This study focuses in particular on the anthropogenic PM_{2.5} loading and its response to the future warming. In this study, the large ensemble simulations from the Community Earth System Model Version 1 (CESM1) throughout the 21st century that are induced by increasing greenhouse gas (GHG) emissions along the trajectory RCP8.5 but retaining the emissions of aerosols and/or their precursors fixed at the year of 2005 level (RCP8.5_FixAerosol2005; Xu and Lamarque, 2018) will be utilized.

2 Data and methods

2.1 PM_{2.5} observational datasets

Surface hourly PM_{2.5} concentration data released since 2013 are taken from the website of the Ministry of Environmental Protection (<http://beijingair.sinaapp.com>, last access: 3 January 2019), which covers 1602 sites across China. The duration of available datasets varies across sites because of the gradual development of the monitoring network in recent years. In our study region of eastern China (east to 100° E), there are 1263 sites remaining after the sites with missing values were removed during 2015–2017. Additionally, surface daily PM_{2.5} concentrations for the Beijing, Shanghai, Guangzhou, and Chengdu cities that had relatively longer monitoring times are also collected from the US Embassy Beijing (<http://www.stateair.net/web/historical/1/1.html>, last access: 3 January 2019).

2.2 CESM1 simulations

The CESM1 is an Earth system model involving the atmosphere, land, ocean, and sea-ice components with a nominal 1° by 1° horizontal resolution (Hurrell et al., 2013). The RCP8.5_FixAerosol2005 simulations are forced by the RCP8.5 scenario, but all emissions of sulfate (SO₄), black carbon (BC), primary organic matter (POM), secondary organic aerosols (SOA; or their precursors), and atmospheric oxidants are fixed at the present-day level (2005). These simulations include 16 ensemble members, differing solely in their atmospheric initial conditions with a tiny random temperature difference (order of 10⁻¹⁴ °C; Kay et al., 2015). For comparison, the CESM1 large ensemble consists of 35-member simulations, forced by the RCP8.5 scenario, that are also employed here. Using these relatively large ensembles can substantially reduce the contribution of the natural variability of the climate system to the result estimation (Xu and Lamarque, 2018).

For the aerosol emission in the RCP scenarios database, just its decadal change is considered rather than the emission in a single year (Lamarque et al., 2011). Here, the years of 2006–2015 are considered as the reference period in the RCP8.5_FixAerosol2005 simulations. The differences of the mean climates from the reference period are largely due to the increase in GHG emissions and are not attributed to the decline in aerosol emissions, as specified in RCP8.5. The changes in anthropogenic PM_{2.5} loadings and anthropogenic air pollution days in our study are thus only a result of the GHG-induced climate change, rather than changes in aerosol emission. Note that just four species of PM_{2.5} components that show a substantial threat to public health are considered here for analysis, including SO₄, BC, POM, and SOA from the CESM1 simulations.

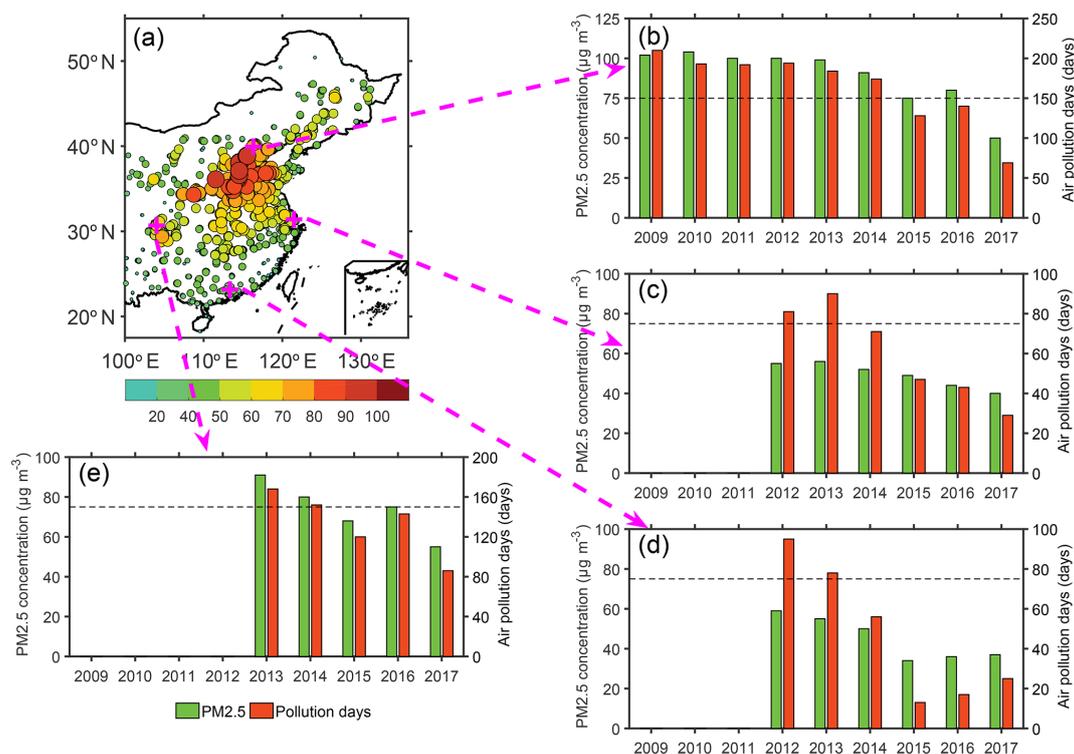


Figure 1. Observed $\text{PM}_{2.5}$ pollution conditions over eastern China during the past years. **(a)** Annual averaged $\text{PM}_{2.5}$ concentration ($\mu\text{g m}^{-3}$) for the years of 2015–2017. **(b)** Variations in annual averaged $\text{PM}_{2.5}$ concentration (green bars) in Beijing city and the corresponding number of the severe $\text{PM}_{2.5}$ pollution days (red bars). The severe pollution days are defined as the daily averaged $\text{PM}_{2.5}$ concentration exceeding $75 \mu\text{g m}^{-3}$. Panels **(c)**, **(d)**, and **(e)** are similar to **(b)**, but for the results of Shanghai, Guangzhou, and Chengdu city, respectively.

2.3 Definition of the fraction of attributable risk

The influences of the GHG-induced climate changes on the anthropogenic air pollution in China are investigated using the metric of the fraction of attributable risk (FAR), which has been widely used for attribute analyses of extreme climate changes (Chen and Sun, 2017; Stott et al., 2004). FAR is defined as $1 - P_0/P_1$, where P_0 is the probability of exceeding a certain threshold during the reference period and P_1 is the probability of exceeding the same threshold during a given period. FAR thus presents the quantitative estimations of effects of the GHG-induced climate changes on the anthropogenic air pollution.

2.4 Definition of stagnation days

The changes in the stagnation days that were induced by the increase in GHG emissions are also evaluated in our study to explore the possible impact of climate change on the anthropogenic air pollution. The day is considered to be stagnant when the daily mean near-surface wind speed is less than 3.2 m s^{-1} , the daily mean 500 hPa wind speed is less than 13 m s^{-1} , and the daily accumulated precipitation is less than 1 mm (Horton et al., 2012). Earlier studies suggested that this air stagnation definition might not be applicable for

China to represent the air pollution condition under the seasonal scales (Feng et al., 2018; Wang et al., 2018). However, the annual mean stagnation generally presents good agreement with that of air pollution across China (Huang et al., 2017, 2018). The changes in the annual mean states of air stagnation over China at the end of 21st century will thus be discussed in the following.

3 Results

3.1 Observational changes in $\text{PM}_{2.5}$ pollution

The days of severe haze pollution increased over the past several decades across eastern China, particularly for the episodes of January 2013, December 2015, and December 2016, when several severe haze alerts were reached. High $\text{PM}_{2.5}$ loading was centralized over the Jing–Jin–Ji (JJJ) region, Shangdong, and Henan provinces, as well as the Sichuan Basin (SCB, Fig. 1a). The annual mean $\text{PM}_{2.5}$ mass concentrations for most sites over these regions exceeded $75 \mu\text{g m}^{-3}$. According to the statistics, at approximately 95 % of sites the annual mean $\text{PM}_{2.5}$ concentration exceeded the WHO recommended 24 h standard ($25 \mu\text{g m}^{-3}$) across eastern China, and there are 65 sites centralized by Beijing where the annual mean $\text{PM}_{2.5}$ concentration was larger than

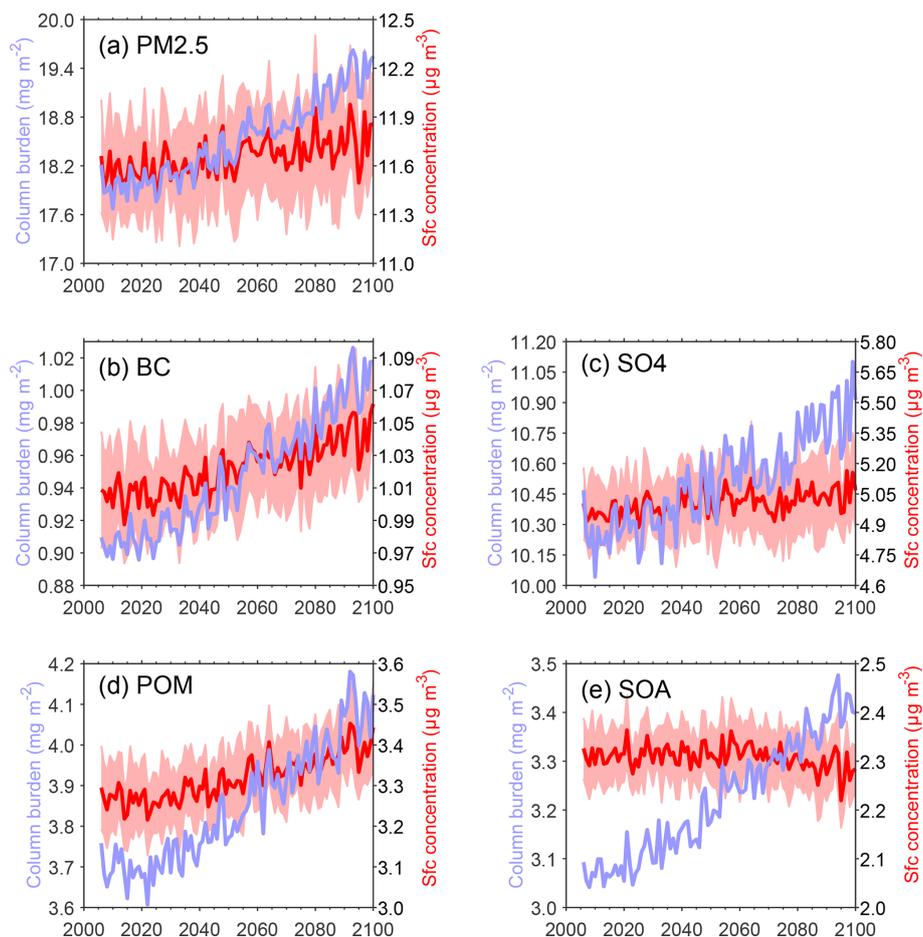


Figure 2. Plots of future changes in the total $\text{PM}_{2.5}$ as well as its associated species averaged over eastern China in terms of the surface concentration ($\mu\text{g m}^{-3}$, right axis in red) and column burden (mg m^{-2} , left axis in blue) from the simulations of the RCP8.5_FixAerosol2005 experiment. (a) $\text{PM}_{2.5}$, (b) BC, (c) SO_4 , (d) POM, and (e) SOA. Ensemble variance (1σ) for surface concentration is shown in red shading.

$75 \mu\text{g m}^{-3}$, which would present the possibility of exposing people to serious health hazards (World Health Organization, 2014).

Regarding the four economic zones of Beijing, Shanghai, Guangzhou, and Chengdu cities in China, serious $\text{PM}_{2.5}$ pollution was reported for recent years, especially for the Beijing and Chengdu regions (Fig. 1). Taking Beijing as an example, the annual mean $\text{PM}_{2.5}$ concentration was stably exceeding $100 \mu\text{g m}^{-3}$, and before 2013 severe air pollution was experienced for more than half the year ($> 75 \mu\text{g m}^{-3}$). Since 2013, China's State Council released its Air Pollution Prevention and Control Action Plan, which requires the key regions, including the JJJ, the Yangtze River Delta (YRD), and the Pearl River Delta (PRD) to reduce their atmospheric levels of $\text{PM}_{2.5}$ by 25 %, 20 %, and 15 %, respectively, by the end of year 2017 (State Council, 2013). Considerable effort was made, and the $\text{PM}_{2.5}$ loading and the air pollution days presented sharp decreases in recent years. However, the strict emission policies have substantial costs for the economic development, which cannot meet the current requirement of the

rapid development of China. Thus, scientifically quantifying the roles of anthropogenic emissions and climate changes holds great importance for seeking the balance between socioeconomic development and emission reduction.

3.2 Simulated changes in anthropogenic $\text{PM}_{2.5}$ pollution

A strong spatial correlation (0.69) is found for the annual mean $\text{PM}_{2.5}$ concentration between the site observation and median ensemble of CESM1 simulations over eastern China (Fig. S1). The high concentrations across eastern China, including the regions centralized by Beijing and Chengdu, are reasonably reproduced. However, a negative bias is obvious. Earlier studies (Li et al., 2016; Yang et al., 2017b, c) have documented that this low bias of aerosol concentration simulated by models is much more complicated in China and the causes mainly involve the uncertainties from aerosol emission amount, emission injection height, lack of nitrate, and

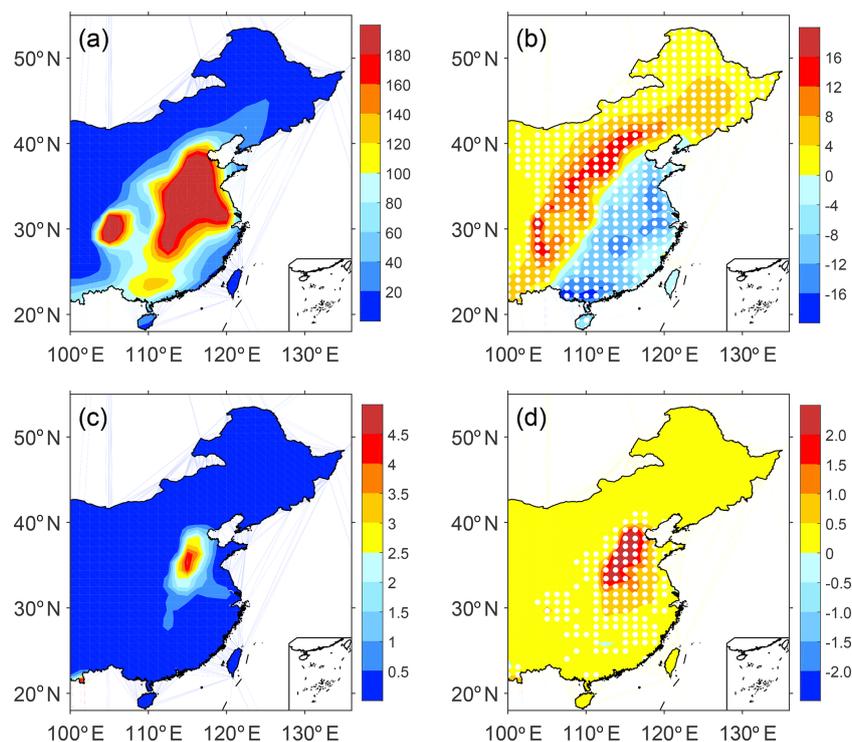


Figure 3. Changes in the anthropogenic PM_{2.5} pollution days across eastern China from the RCP8.5_FixAerosol2005 experiment. The top panels (a, b) show the changes in light air pollution days ($> 25 \mu\text{g m}^{-3}$) and the bottom panels (c, d) show the results of severe air pollution days ($> 75 \mu\text{g m}^{-3}$). The left panels (a, c) illustrate the annual averaged severe pollution days in 2006–2015 and the right panels (b, d) show changes in the pollution days at the end of the 21st century compared to 2006–2015. Dots in (b) and (d) mean the changes are significant at the 95 % confidence level using Student *t* test for all years and ensembles. Units: days.

aerosol treatment in the model as well as the coarse model resolution.

The median ensemble-mean change in the PM_{2.5} surface concentration presents strong regional dependence across China, with significantly decreasing trends over the southeastern part of eastern China and significantly increasing trends over the other regions throughout the 21st century (Fig. S2), even though the emissions are constant throughout the experiment. The regional differences in the total PM_{2.5} changes are mainly due to SO₄, which can account for approximately 50 % of the total PM_{2.5} mass (Xu and Lamarque, 2018). The species of BC and POM are reported to significantly increase in the 21st century across eastern China, although the aerosol emissions were fixed at the level in 2005. Figure 2 presents the simulated PM_{2.5} loadings from the CESM1, in terms of column burden and surface concentration, that are significantly increasing throughout the 21st century. The increase in the total PM_{2.5} is approximately 8 % for the column burden and 2 % for the surface concentration at the end of the 21st century (2090–2099) compared to the current state (2006–2015). These increasing trends of PM_{2.5} loadings are mainly due to the significant increases in the major PM_{2.5} species, except for SOA, in which the surface concentration presents a slight decrease. Furthermore, the in-

creases in all major PM_{2.5} species in terms of column burden (BC: 11 %, SO₄: 6 %, SOA: 11 %, and POM: 11 %) show stronger than the surface concentration (BC: 4 %, SO₄: 2 %, SOA: −1 %, and POM: 4 %).

For comparison, we also evaluated the future changes in PM_{2.5} concentrations and the associated species along the RCP8.5 forcing trajectory from the large ensemble simulations of CESM1 (figure not shown). Different from changes in aerosol concentrations under the fixed aerosol simulations, the PM_{2.5} concentrations and the associated species present uniformly decreasing trends across eastern China from the simulations along the RCP8.5 forcing. The decreasing trends in the RCP8.5 simulations are mainly attributed to the prescribed decrease in aerosol forcing in the future in the RCP database (Xu and Lin, 2017). The climate change induced by the GHG-warming might exacerbate the air pollution, but the impacts cannot compensate the prescribed decreasing trend of aerosol concentration.

As mentioned above, the PM_{2.5} surface concentration in the two economic zones of YRD and PRD present a negative response to the GHG-induced warming, while the corresponding column burden shows significantly increasing trends (Fig. S3). The decreases in the surface concentration over these two zones are primarily contributed by the

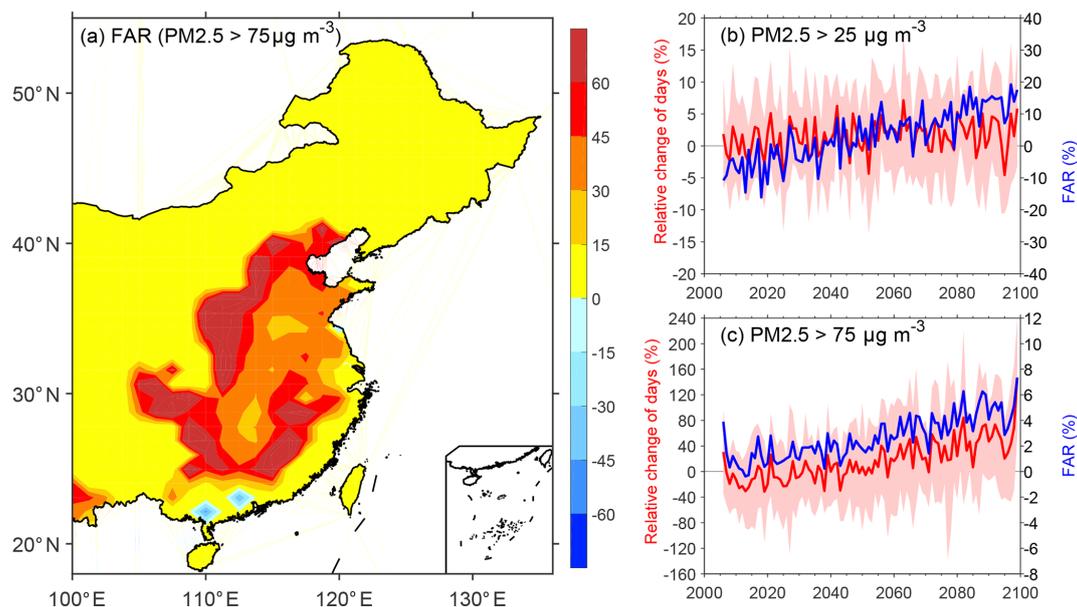


Figure 4. Attributable changes in anthropogenic air pollution days to the increased greenhouse gases emissions. **(a)** Spatial distribution of FAR for the changes in severe PM_{2.5} pollution (> 75 µg m⁻³) at the end of the 21st century over eastern China. **(b)** Regional averaged relative changes in air pollution days (left axis in red; > 25 µg m⁻³) and the corresponding variation in FAR (right axis in blue). Ensemble variance (1σ) for the relative changes in pollution days is shown in red shading. **(c)** is similar to **(b)**, but for the severe PM_{2.5} pollution days. Units: percentage (%).

changes in SO₄ and SOA, while there are no obvious trends for BC and POM (Figs. S4–S7). The robust response of the increased surface wind speed and decreased upper-level wind speed to GHG warming can be partly responsible for the changes in the major PM_{2.5} species in these two zones, which will be further discussed. Over the zones of JJJ and SCB, both the PM_{2.5} concentrations and the associated major PM_{2.5} species present the significantly rising trends throughout the 21st century. For the surface concentration, PM_{2.5} is reported to increase by 3 % and 4 % in the regions of JJJ and SCB, respectively, at the end of the 21st century. The BC is reported to increase by 4 % and 8 % for JJJ and SCB, respectively. The other species, such as SO₄ and POM, increase by 4 % and 4 %, respectively, in the JJJ regions and by 2 % and 9 %, respectively, in SCB regions. Relatively stronger responses can be seen in changes in the column burden for all major species (Figs. S4–S7). The increased concentrations of PM_{2.5} species finally result in significantly increasing trends of the total PM_{2.5} loading over these two regions, which will present a more direct effect on human health.

The increase in PM_{2.5} surface concentration throughout the 21st century substantially leads to the significant increase in the light anthropogenic PM_{2.5} pollution days (PM_{2.5} > 25 µg m⁻³) across the northwestern part of eastern China (Fig. 3). Due to the decrease in PM_{2.5} concentration over the southeastern part of eastern China, the light anthropogenic air pollution days can be expected to decrease in this region. Estimation shows that the number of the light air pol-

lution days would be decreased by approximately 10 days at the end of the 21st century compared to the early period of this century in the region. However, the annual mean light air pollution days are reported to increase on average over eastern China at the end of this century despite the aerosol emission being constant throughout the experiment. In contrast to the light air pollution days, the severe anthropogenic air pollution days (PM_{2.5} > 75 µg m⁻³) show a positive response to the GHG-induced warming across eastern China, particularly for the regions around JJJ in which the high PM_{2.5} concentration was localized (Fig. 3). The severe air pollution days are estimated to increase over this region by more than 2 days at the end of this century when compared to the early period. Considering the underestimation in aerosol concentration by CESM1 in China, the percentile threshold metric is also applied here to estimate the future changes in light (90th) and severe (99th) air pollution days. Similar results can be obtained (Fig. S8).

3.3 Attributable changes due to the GHG warming

Although the aerosol emission was constant throughout the experiment, our study reveals that the PM_{2.5} loadings and their associated pollution days still present increases throughout the 21st century, primarily resulting from the impact of climate change induced by the GHG warming. One may ask how large a contribution the climate change exerts on the changes in anthropogenic air pollution. To quantitatively address this issue, the framework of the “fraction of

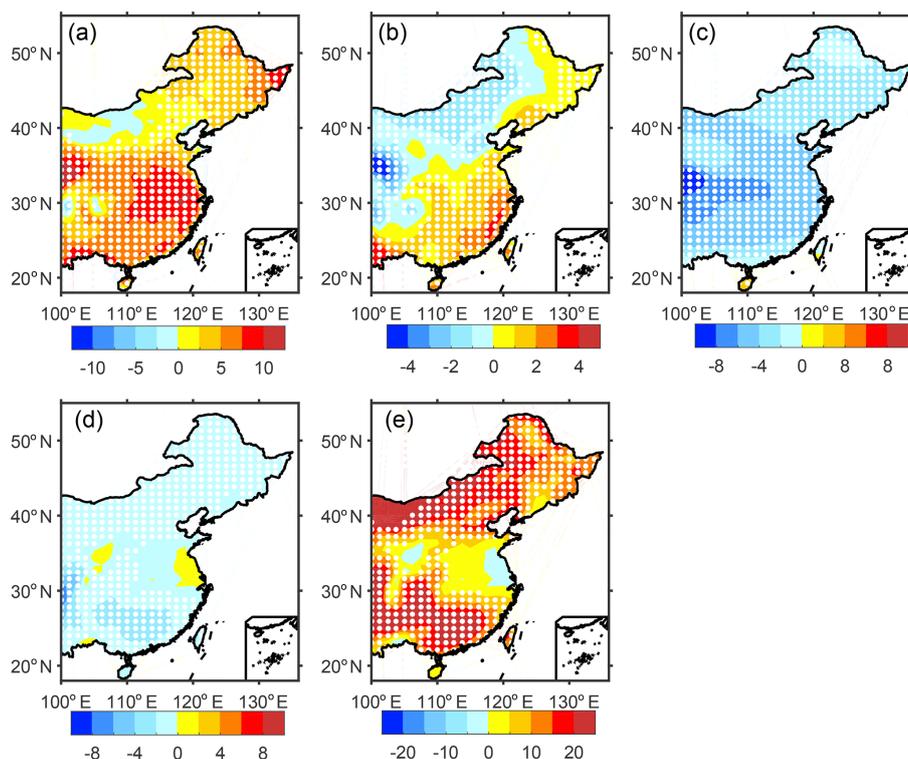


Figure 5. Simulated changes in weather conditions of the air pollution across eastern China due to the GHG-induced warming. (a) Changes in the planetary boundary layer height (PBLH) at the end of the 21st century compared to the years of 2006–2015 from the RCP8.5_FixAerosol2005 experiment. Panels (b) and (c) are similar to (a) but for the wind speed at near-surface and 500 hPa levels, respectively. (d) Changes in the light precipitation days (daily accumulated precipitation < 10 mm) at the end of the 21st century compared to the current state. Panel (e) is similar to (d) but for the heavy precipitation days (> 10 mm). Dots in the figure mean the changes are significant at the 95 % confidence level using Student *t* test for all years and ensembles. Units: percentage (%).

attributable risk” that has been widely used for attribute analyses of climate extreme changes (Chen and Sun, 2017; Stott et al., 2004) is employed in this study.

Figure 4 shows the percentage changes in the anthropogenic air pollution days throughout the 21st century over eastern China and their associated FAR variations. The regional averaged anthropogenic air pollution days present an obvious increase in the 21st century as addressed above. Correspondingly, synchronous increasing trends can be found in FAR for both light and severe anthropogenic air pollution days. For the light pollution days, FAR is estimated to be 28 % at the end of the 21st century, implying that approximately 28 % of the pollution days are contributed by the climate change that was induced by the GHG warming. For the severe pollution days, FAR shows a relatively smaller value of approximately 11 %. Furthermore, the high FAR values are mainly located over the regions of high PM_{2.5} loadings concentrated over eastern China, suggesting considerably stronger effects of climate change in these regions. Note that the FAR values estimated in this research may be underestimated because the GHG-induced warming impact was involved in the selected reference period that resulted in the

overestimation of the probability of anthropogenic air pollution days.

3.4 Effects of the changes in meteorological conditions

We further examined the changes in meteorological conditions induced by the GHG warming that exerted effects on air pollution. Our results show that the local boundary layer height presents as higher under the warming scenario (Fig. 5a), which benefits the vertical transport of the air pollutant.

However, a robust negative response of the horizontal advection to the GHG-induced warming across eastern China can be found in the troposphere (Fig. 5b, c), facilitating air pollutant accumulation. The change in surface wind speed in response to the GHG warming is highly similar with the variation in PM_{2.5} surface concentration, with wind speed increasing in the southeastern part of eastern China and decreasing in the northwestern part. Variations in surface wind speeds are thus mainly responsible for the changes in PM_{2.5} surface concentration over eastern China. Different responses can be found for the tropospheric upper-level wind speeds, which are reported to substantially decrease.

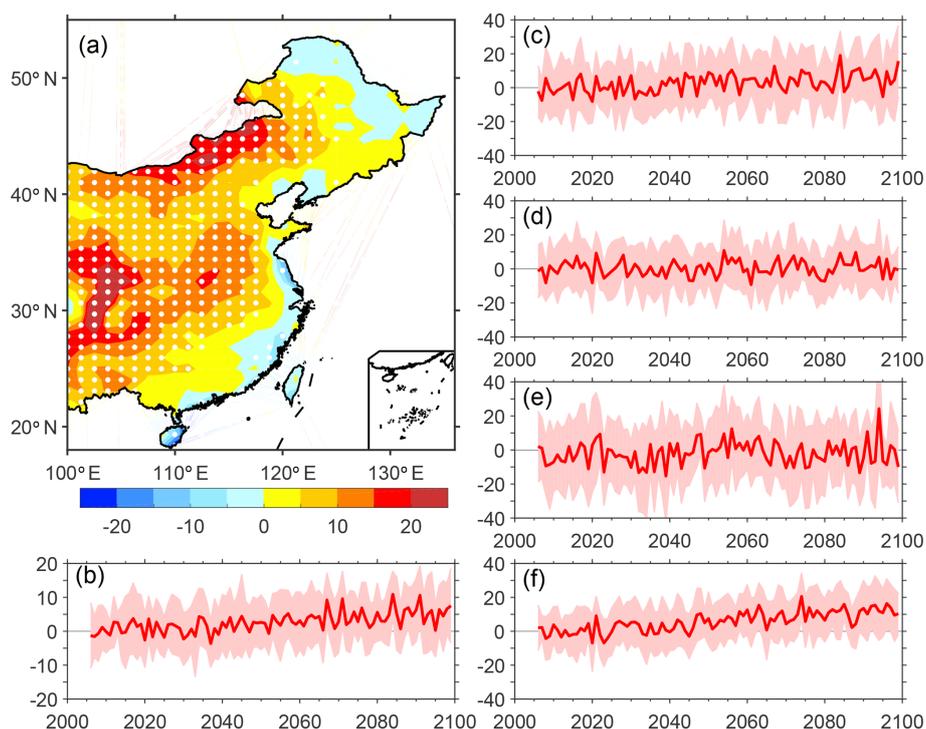


Figure 6. Changes in the stagnant conditions across China due to the GHG-induced warming. **(a)** Distribution of the relative changes in the stagnation days at the end of the 21st century against the current state (2006–2015). Dots mean the changes are significant at the 95 % confidence level using Student t test for all years and ensembles. **(b)** Variations in the regional averaged stagnation days over eastern China. Ensemble variance (1σ) is shown in red shading. Panels **(c)**, **(d)**, **(e)**, and **(f)** are similar to **(b)**, but for the results of four Chinese economic zones, i.e. JJJ, YRD, PRD, and SCB. Units: percentage (%).

These decreases would directly result in significant increases in the stagnation days over eastern China, particularly over the northern region and SCB (Fig. 6). The decreasing trend of wind speed in the 21st century across China not only exists in CESM1, but also happens in the other global climate models that participated in Coupled Model Intercomparison Project Phase 3 (CMIP3) and CMIP5 (Jiang et al., 2010a; McInnes et al., 2011), as well as in regional climate models (Jiang et al., 2010b).

In response to the GHG-induced warming, the stagnation days over eastern China are estimated to increase by 6 % at the end of 21st century compared to the current period. For the specific economic zones, the stagnation days over the SCB and JJJ regions show considerably stronger rising trends, while relatively weaker increases are observed over the YRD and PRD regions. The number of stagnation days is estimated to increase by 13 % and 6 % at the end of the 21st century for the SCB and JJJ regions, respectively. Briefly, though the atmospheric stratification appears to be considerably more unstable in response to the GHG warming, the weakened horizontal advection would substantially increase the stagnation days over eastern China, which provides a beneficial background for the air pollutant accumulation and further increases the occurrence probability of the anthropogenic air pollution events.

Earlier studies have documented a significant increase in total precipitation across China due to the GHG-induced warming (Chen, 2013; Li et al., 2018; Wang et al., 2012), which seems to represent a conflict with the increase in the anthropogenic air pollution days. To resolve this issue, the precipitation changes in terms of light precipitation days (daily accumulated precipitation < 10 mm) and heavy precipitation days (> 10 mm) are further examined (Fig. 5d, e). Clearly, the heavy precipitation days present an increase, while the light precipitation days show a decrease, across eastern China in response to the warming. Though the precipitation shifts toward heavy precipitation events, its cleansing impact on air pollutants has not increased because an increase in heavy precipitation days appears to be insufficient to further enhance the wet removal ability (Xu and Lamarque, 2018). In contrast, the decrease in light precipitation days substantially weakens the wet deposition of air pollutants, leading to the increase in the $PM_{2.5}$ loading, as well as anthropogenic air pollution days. The future changes in precipitation days presented here are robust. Both the increasing trends of heavy precipitation days and the decreasing trends of light precipitation days are also obvious across China, simulated by the CMIP5 models (Chen and Sun, 2013, 2018) as well as the regional climate models (Gao et al., 2012).

4 Conclusions

The world is projected to experience increased disasters, such as heat waves, flash floods, and storms, due to the continuous global warming induced by the GHG increase. The research question we aim to address in this study is how the GHG warming would affect the anthropogenic PM_{2.5} pollution across China. Our evaluations show that the anthropogenic PM_{2.5} loadings, as well as the anthropogenic PM_{2.5} pollution days, would increase under the global warming conditions, even the aerosol emissions fixed at current levels. More stringent regulations are thus suggested for regional aerosol emissions to maintain the air quality standard as the current state.

The climate changes induced by the GHG warming exert their effects on the anthropogenic air pollution across eastern China via two ways that are of interest in this study. First, the weakened tropospheric wind speed induced by the GHG warming would result in a decrease in the horizontal advection and lead to an increase in the number of stagnation days, facilitating the local accumulation of air pollutants. Second, the number of light precipitation days would decrease due to the GHG-induced warming, although the total precipitation would clearly increase across China. This shift toward more no-rainfall days would further weaken the wet deposition of PM_{2.5} pollutants. Thus, the increased stagnation days and decreased light precipitation days provide a beneficial background for the occurrence of anthropogenic air pollution. Of course, under the warming scenarios, a large discrepancy exists among the different meteorological processes that benefit the air pollution at the current state, leading to the fuzzy recognition of air pollution change. For example, the boundary layer height shows an increase in response to the GHG warming that may strengthen the vertical dissipation of air pollutants. Thus, more studies are suggested in the future to further understand the mechanisms governing air quality across China.

Data availability. All the datasets that were used for the analyses in this study are available for the public to download. The dataset of the site-observed surface hourly PM_{2.5} concentration across China is available at <http://beijingair.sinaapp.com/> (last access: 3 January 2019, China National Environmental Monitoring Centre, 2019). The surface daily PM_{2.5} concentrations for the Beijing, Shanghai, Guangzhou, and Chengdu cities that had relatively longer monitoring times can be downloaded from the website <http://www.stateair.net/web/historical/1/1.html> (last access: 3 January 2019, US Embassy Beijing, 2019). The CESM1 RCP8.5_FixAerosol2005 simulations are available at <https://www.earthsystemgrid.org/dataset/ucar.cgd.cesm4.output.html> (last access: 3 January 2019, Xu et al., 2018).

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Competing interests. The authors declare that they have no conflict of interest.

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